



# Alkali-resistant NO<sub>x</sub> reduction over FeVO<sub>4</sub>/TiO<sub>2</sub> catalysts *via* regulating the electron transfer between Fe and V

Yunang Dong<sup>1</sup>, Penglu Wang<sup>1</sup>, Xiangyu Liu, Jiang Deng, Aling Chen, Lupeng Han, Dongsong Zhang\*

International Joint Laboratory of Catalytic Chemistry, State Key Laboratory of Advanced Special Steel, College of Sciences, Shanghai University, Shanghai 200444, China

## ARTICLE INFO

### Article history:

Received 7 February 2023

Revised 5 May 2023

Accepted 30 May 2023

Available online 2 June 2023

### Keywords:

Air pollution control

NO<sub>x</sub> reduction

Selective catalytic reduction

Alkali metals

Electron transfer

## ABSTRACT

The presence of alkali metals in exhaust gas from stationary resources causes a grand challenge for the practical application of selective catalytic reduction (SCR) of NO<sub>x</sub> with NH<sub>3</sub>. Here, alkali-resistant NO<sub>x</sub> reduction has been successfully implemented *via* tailoring the electron transfer over Fe and V species on FeVO<sub>4</sub>/TiO<sub>2</sub> catalysts. The strong interaction between Fe and V induced electron transfer from V to Fe and strengthened the adsorption and activation of NH<sub>3</sub> and NO over active VO<sub>x</sub> sites. In the presence of K<sub>2</sub>O, the strong electron withdrawing effect of Fe offset the electron donating effect of K on the VO<sub>x</sub> species, thus protecting the active species VO<sub>x</sub> to maintain the NO<sub>x</sub> reduction ability. The enhanced adsorption and activation of NH<sub>3</sub> allowed SCR reaction to proceed *via* E-R mechanism even after K<sub>2</sub>O poisoning. This work elucidated the electronic effects on the alkali metals resistance of traditional ferric vanadate SCR catalysts and provided a promising strategy to design SCR catalysts with superior alkali resistance.

© 2023 Published by Elsevier B.V. on behalf of Chinese Chemical Society and Institute of Materia Medica, Chinese Academy of Medical Sciences.

The emissions of nitrogen oxides (NO<sub>x</sub>) from coal-fired power stations and steel plants can trigger a series of environmental issues [1–3]. Selective catalytic reduction of NO<sub>x</sub> with NH<sub>3</sub> (NH<sub>3</sub>-SCR) is the most effective technology to eliminate NO<sub>x</sub> currently [4,5]. Due to the satisfied SCR performance, V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>(MoO<sub>3</sub>)/TiO<sub>2</sub> catalysts have been widely applied in stationary sources [6–8]. However, there still exist some drawbacks, such as the narrow operating temperature window (300–400 °C) and the over-oxidation of SO<sub>2</sub> [9,10]. Furthermore, the presence of alkali metals in the exhaust gas can lead to severe deactivation of the catalysts [11–13]. Therefore, it is quite urgent to enhance the activity and alkali resistance of V-based catalysts to broaden the application among NH<sub>3</sub>-SCR [14–16].

So far, the FeVO<sub>4</sub> based catalyst has aroused extensive research interests because of its excellent low temperature activity and SO<sub>2</sub> resistance [17–19]. Liu *et al.* proposed that the well-dispersed FeVO<sub>4</sub> catalysts exhibited high activity and N<sub>2</sub> selectivity. They also revealed that the true active sites in FeVO<sub>4</sub> catalysts were VO<sub>x</sub> species [20]. Mu *et al.* constructed a wide range of Fe<sub>1-x</sub>V<sub>x</sub>O<sub>8</sub> catalysts with various Fe to V ratios and found that the electron-induced effects between Fe and V together with the interaction

between FeVO<sub>4</sub> and Fe<sub>2</sub>O<sub>3</sub> species would greatly enhance the SCR performance [21]. However, after K<sub>2</sub>O poisoning, the activity of FeVO<sub>4</sub> would decreased dramatically because of the destruction of acidity and redox properties [22,23]. Traditional strategies for improving the resistance to alkali metals could be divided into two routes: (1) Enhancing the surface acidity and (2) separating the active sites and the poisoning sites [24–27]. Basically, alkali metals are kinds of electron donors and their poisoning effects on catalysts mainly rely on the electron impact *via* reducing the valence of the bonded active metal elements, thus undermining the Lewis acidity and redox capacity of the catalysts. Hence, weakening the electronic effect of alkali metals on the active species plays a crucial role in the resistance promotion of alkali metals on SCR catalysts [28,29]. Li *et al.* modified commercial V<sub>2</sub>O<sub>5</sub>-WO<sub>3</sub>/TiO<sub>2</sub> catalysts with Ce and Cu and verified that the interaction between V, Ce, and Cu increased the ratio of V<sup>5+</sup>/(V<sup>5+</sup>+V<sup>4+</sup>), which contributed to the excellent K<sub>2</sub>O resistance of the catalysts [30]. However, the effect of electronic interactions between Fe and V on the alkali resistance of catalysts as well as the specific electron effect of alkali metals on the active V species had still not been clearly defined.

In this work, the FeVO<sub>4</sub> based catalysts with strong or weak interactions between Fe and V have been fabricated to investigate the role of electronic effects on the alkali resistance of SCR catalysts. The FeVO<sub>4</sub> catalyst prepared *via* impregnation method (i-

\* Corresponding author.

E-mail address: dszhang@shu.edu.cn (D. Zhang).

<sup>1</sup> These authors contributed equally to this work.

FeV/Ti) is characterized as tightly bounded Fe and V with strong interaction, which shows superior SCR performance and alkali resistance than the catalyst with weaker interaction between Fe and V via co-precipitation preparation approach (denoted as p-FeV/Ti). The strong electronic withdraw ability of Fe increases the content of  $V^{5+}$  on the surface, thus leading to promoted Lewis acidity and redox ability, which is favorable for the adsorption and activation of  $NH_3$  and NO. After  $K_2O$  poisoning, the electron-absorbing effects of Fe have offset the electron-donating effects of  $K_2O$  on active V sites, therefore protecting the active  $VO_x$  species to express the superior catalytic performance of  $NO_x$  reduction. Subsequently, the adsorbed and activated  $NH_3$  on the surface could react with gaseous NO and  $O_2$  following Eley-Rideal mechanism even in the presence of  $K_2O$ . This work provides a deep understanding of the effects of electronic interactions on the alkali resistance of  $FeVO_4$  catalysts and paves a promising way to further design SCR catalysts with improved alkali resistance.

The i-FeV/Ti catalyst was prepared through a traditional impregnation method. First, 0.3000 g oxalic acid was dissolved in 50 mL deionized water, then 0.0684 g ammonium metavanadate was added and stirred for 0.5 h, followed by the addition of 0.2363 g  $Fe(NO_3)_3 \cdot 9H_2O$  and 1 g  $TiO_2$ . After stirring for another 2 h, they were dried using a rotary evaporator at 50 °C and then calcined at 500 °C for 4 h with a ramping rate of 5 °C/min. The resulting samples were labeled as i-FeV/Ti. The p-FeV/Ti catalyst was prepared by a co-precipitation method and alkali poisoned samples were obtained by impregnating 1 wt%  $K_2O$  onto fresh catalysts. All details can be found in the Supporting information.

SCR performance was tested through a fixed-bed quartz flow reactor with 0.3 g catalysts. The reactant gas was composed of 500 ppm NO, 500 ppm  $NH_3$ , 5 vol%  $O_2$  and the carrier gas  $N_2$ . The total gas hourly space velocity (GHSV) was controlled to be 50,000  $h^{-1}$ .  $NO_x$  conversion and  $N_2$  selectivity were calculated by the following equations (Eqs. 1 and 2):

$$NO_x \text{ conversion (\%)} = \frac{[NO_x]_{in} - [NO_x]_{out}}{[NO_x]_{in}} \times 100\% \quad (1)$$

$$N_2 \text{ selectivity (\%)} = \left( 1 - \frac{2[N_2O]_{out}}{[NH_3]_{in} + [NO_x]_{in} - [NH_3]_{out} - [NO_x]_{out}} \right) \times 100\% \quad (2)$$

$[NO_x]$  represented the total concentration of NO and  $NO_2$ , while  $[NO_x]_{in}$ ,  $[NO_x]_{out}$ ,  $[NH_3]_{in}$ ,  $[NH_3]_{out}$ ,  $[N_2O]_{out}$  indicated the concentrations of the corresponding gas at the inlet and outlet, respectively.

More detailed experimental parameters and characterizations could be found in Supporting information.

The i-FeV/Ti catalyst was prepared by impregnation method, which was characterized as tightly bounded Fe and V. Fig. 1 shows the  $NO_x$  conversion and the  $N_2$  selectivity of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti, and K-p-FeV/Ti catalysts under a gaseous hourly space velocity (GHSV) of 50,000  $h^{-1}$ . The i-FeV/Ti catalyst exhibits a broad temperature window of 210–390 °C with  $NO_x$  conversion over 90% and  $N_2$  selectivity nearly 100%. By contrast, the  $NO_x$  conversion of p-FeV/Ti is lower within 150–300 °C. After  $K_2O$  poisoning, two catalysts show quite different resistance performance. The SCR performance of i-FeV/Ti declines slightly within the whole temperature region. In contrast, the  $NO_x$  conversion of p-FeV/Ti decreases greatly from 96% to 31% at 300 °C and stays below 45% within 120–390 °C. We also adjusted the content of Fe and V in i-FeV/Ti and found that when the ratio of Fe to V was 1:1, the catalyst showed the best activity and alkali resistance (Fig. S1 in Supporting information). In addition, we also test the SCR performance of i-FeV/Ti and p-FeV/Ti in the copresence of 1% K and 3% Pb and found that the performance of K&Pb-i-FeV/Ti remains superior

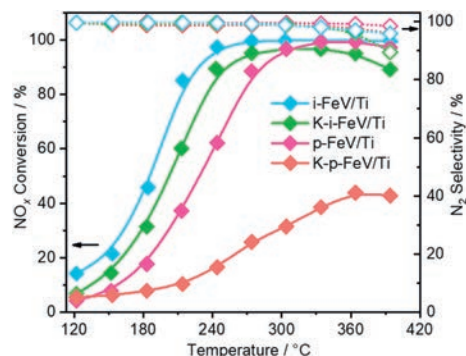
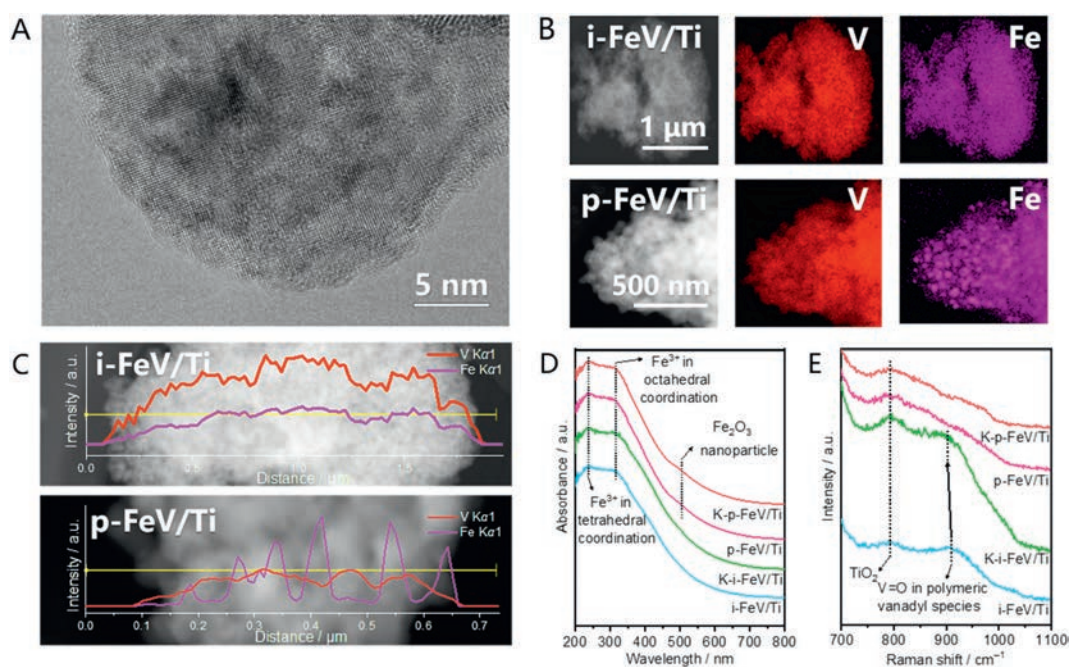


Fig. 1.  $NO_x$  conversion and  $N_2$  selectivity of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti and K-p-FeV/Ti catalysts. Reaction conditions: 500 ppm of NO, 500 ppm of  $NH_3$ , 5 vol%  $O_2$ ,  $N_2$  as the balance gas, GHSV of 50,000  $h^{-1}$ .

than that of K&Pb-p-FeV/Ti (Fig. S2 in Supporting information). Besides, the activation energy of i-FeV/Ti and p-FeV/Ti was also tested under high GHSV (200,000  $h^{-1}$ ) with  $NO_x$  conversion below 20% (Figs. S3 and S4 in Supporting information). It was indicated that the i-FeV/Ti catalyst shows lower activation energy (39.37 kJ/mol) than p-FeV/Ti (51.07 kJ/mol), demonstrating its higher intrinsic activity. Furthermore, after  $K_2O$  poisoning, the activation energy of K-i-FeV/Ti (55.33 kJ/mol) remains lower than that of K-p-FeV/Ti (64.76 kJ/mol). The tightly bounded structure between Fe and V is more conducive to induce strong electronic interactions, that should be the reason why i-FeV/Ti exhibited superior low temperature activity and alkali resistance [20,31].

In order to explore the intrinsic reasons for the different alkali resistance of i-FeV/Ti and p-FeV/Ti catalysts, the structural characteristics of catalysts were investigated. X-ray diffraction (XRD) patterns reveal the same diffraction peaks belonged to anatase and  $FeVO_4$  for all catalysts (Fig. S5 in Supporting information). There are no obvious changes after  $K_2O$  poisoning, indicating that  $K_2O$  makes no obvious effects on crystal phase. Besides, the content of Fe and V in all catalysts was investigated by inductively coupled plasma-atomic emission spectrometry (ICP-AES) (Table S1 in Supporting information). The results suggest that i-FeV/Ti and p-FeV/Ti possess nearly the same amount of Fe and V, well excluding the effects of different amount of active species on the differential alkali resistance. Moreover,  $N_2$  adsorption-desorption experiments were performed to characterize the pore structure of catalysts and the specific surface area were calculated by the Brunauer-Emmett-Teller model. It could be seen that the specific surface area of p-FeV/Ti is slightly larger than that of i-FeV/Ti (Table S2 in Supporting information). After  $K_2O$  poisoning, the specific surface area values for both samples decrease slightly, probably due to the blockage of pore channels [32,33].

To probe the element distribution on the surface of these catalysts, the high-resolution transmission electron microscopy (HRTEM) and energy-dispersive X-ray spectroscopy (EDS) were carried out. The HRTEM image in Fig. 2A shows that there are no large  $FeVO_4$  crystalline particles on the  $TiO_2$  support for i-FeV/Ti, indicating that the active components are more likely to present in an amorphous form. Besides, the Fe and V elements are uniformly dispersed on the i-FeV/Ti catalyst, as seen from the elemental distribution mapping results in Fig. 2B. However, there are some aggregations of Fe species on p-FeV/Ti, exhibiting grain-like characteristics in Fig. 2B. It could also be found that distribution tendencies of Fe and V are almost the same on the surface of i-FeV/Ti in Fig. 2C, indicating that Fe element is closely combined with V species. However, the line sweep results for Fe on p-FeV/Ti catalysts show a wave-like trend in Fig. 2C, further demonstrating the inhomogeneous distribution of Fe. The poor distribution of Fe



**Fig. 2.** (A) HRTEM image of i-FeV/Ti. (B) EDX mapping of i-FeV/Ti and p-FeV/Ti. (C) EDX Line sweep of i-FeV/Ti and p-FeV/Ti. (D) UV-vis spectra and (E) Raman spectra of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti, and K-p-FeV/Ti catalysts.

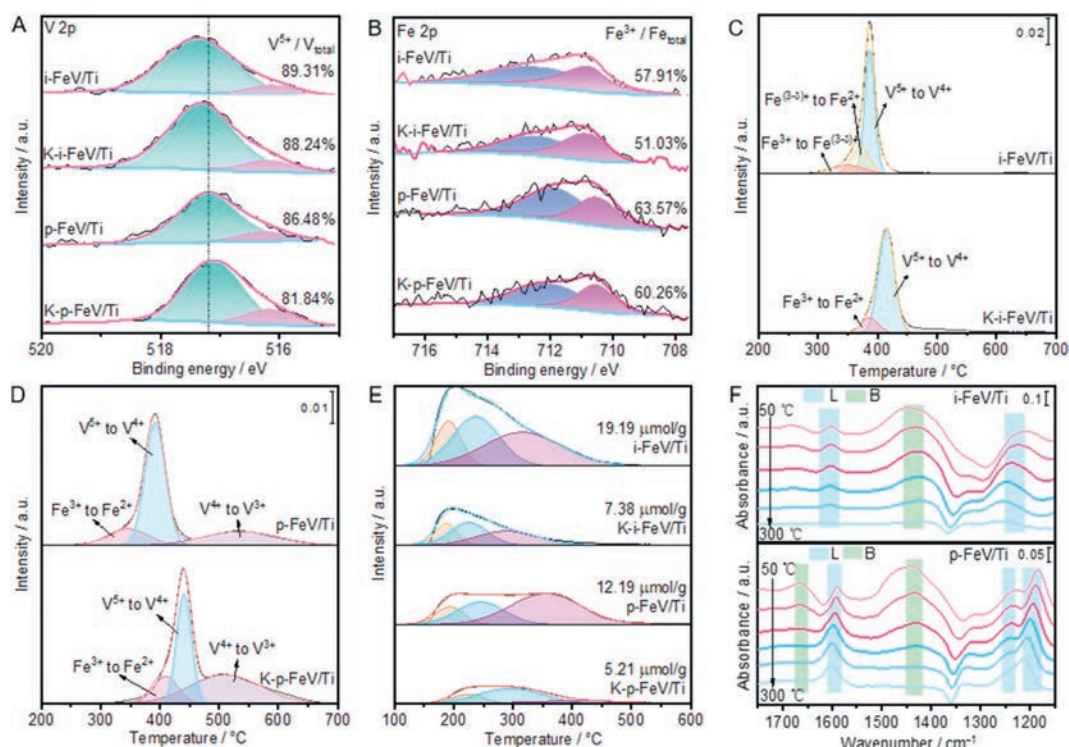
leads to limited bindings with V species, resulting in weaker electron interactions between Fe and V on p-FeV/Ti catalyst, which is the main reason for the worse activity and alkali resistance.

In addition, UV-vis spectra were performed to further characterize the coordination structure of these catalysts. As shown in Fig. 2D, the peaks at 234 and 311 nm are attributed to  $\text{Fe}^{3+} \leftarrow \text{O}$  charge transfer of isolated Fe ions in tetrahedral and octahedral coordination, respectively. While the d-d transition band located at 799 nm is the fingerprint of  $\text{V}^{4+}$  [34,35]. Additionally, p-FeV/Ti exhibits an obvious band at 501 nm which should be ascribed to  $\text{Fe}_2\text{O}_3$  nanoparticles [36,37]. Similar bands are not observed on i-FeV/Ti, which is consistent with the EDX mapping and line sweep results. Moreover, the *in situ* UV-vis DRS spectra were also performed at 240 °C to verify the active species on i-FeV/Ti and p-FeV/Ti (Figs. S6 and S7 in Supporting information). After the introduction of  $\text{NH}_3$  or  $\text{NO}+\text{O}_2$ , the variations mainly occur in the range of 510–800 nm, corresponding to the d-d transition bands of  $\text{V}^{3+}/\text{V}^{4+}$  [38]. As for the bands at 234, 311 and 501 nm corresponding to the Fe sites, there is no obvious change after the flowing of reaction gasses. This result demonstrates that the real active sites in the  $\text{FeVO}_4$  catalysts are polymeric  $\text{VO}_x$  species, which is consistent with the reported literatures [20,39].

Raman spectra were also collected to further identify the structure of all catalysts in Fig. 2E. All the spectra show the second-order feature of  $\text{TiO}_2$  at  $795\text{ cm}^{-1}$  [40]. The band at  $923\text{ cm}^{-1}$  should be ascribed to  $\text{V}=\text{O}$  stretches at polymeric vanadyl species for i-FeV/Ti [6]. As for p-FeV/Ti and K-p-FeV/Ti catalysts, the peaks corresponding to the polymeric V species are significantly weaker. Due to the overlay of  $\text{TiO}_2$  peaks, the bands related to Fe species could not be observed. Therefore, to further exclude the influence of support on the observation of Fe and V species among i-FeV/Ti and p-FeV/Ti catalysts, we continued to prepare bulk i-FeV and p-FeV materials to perform the Raman spectroscopy. It can be observed that the bands at 320, 363, 454, 645, 719, 753, 821, 882, 915,  $945\text{ cm}^{-1}$  should be attributed to the representative  $\text{FeVO}_4$  and bands at 212,  $272\text{ cm}^{-1}$  are ascribed to hematite phase (Fig. S8 in Supporting information) [41]. It is obvious that p-FeV shows much stronger characteristic of  $\text{Fe}_2\text{O}_3$  than i-FeV, and the peaks at

$645\text{ cm}^{-1}$  corresponding to the bridging Fe-O-V bands are more prominent on i-FeV, indicating tighter bonds between Fe and V [42]. The Raman spectra and EDS results demonstrate that i-FeV/Ti catalysts possessed more Fe-O-V sites than p-FeV/Ti catalysts.

To clarify the electronic interactions between Fe and V species, X-ray photoelectron spectroscopy (XPS) was performed to probe the valence states of elements on the catalysts surface. As shown in Fig. 3A, the V  $2p_{3/2}$  characteristic peak of i-FeV/Ti is centered at  $517.39\text{ eV}$  [43]. Compared with p-FeV/Ti ( $517.19\text{ eV}$ ), V 2p peaks of i-FeV/Ti show higher binding energies. This variation can be caused by the transfer of more electrons from V to Fe. After K poisoning, the location of V  $2p_{3/2}$  characteristic peak on K-i-FeV/Ti barely changes but that on K-p-FeV/Ti shifts to the lower binding energy, indicating that the introduction of  $\text{K}_2\text{O}$  reduces the valence states of V element on the surface for p-FeV/Ti but barely effect i-FeV/Ti. Besides, the  $\text{V}^{5+}/(\text{V}^{5+}+\text{V}^{4+})$  ratio of i-FeV/Ti (89.31%) is higher than that of p-FeV/Ti (86.48%), which is in favor of  $\text{NH}_3$  adsorption due to the stronger Lewis acidity [44]. Furthermore, the  $\text{V}^{5+}/(\text{V}^{5+}+\text{V}^{4+})$  ratio slightly changes after  $\text{K}_2\text{O}$  is introduced (88.24%) for i-FeV/Ti. As for p-FeV/Ti, the ratio decreases from 86.48 to 81.84% after  $\text{K}_2\text{O}$  poisoning. This phenomenon indicates that  $\text{K}^+$  imposes less influence on the electronic property of  $\text{VO}_x$  species of i-FeV/Ti [39]. The ratios of  $\text{Fe}^{3+}/(\text{Fe}^{3+}+\text{Fe}^{2+})$  were also calculated based on the Fe 2p XPS results shown in Fig. 3B. There are less  $\text{Fe}^{3+}$  on i-FeV/Ti (57.91%) than on p-FeV/Ti (63.57%). The different ratio of  $\text{Fe}^{3+}$  on two samples exactly proves the electron transfer from V to Fe over i-FeV/Ti, which has mentioned above. When  $\text{K}_2\text{O}$  is introduced, the amount of  $\text{Fe}^{3+}$  declined from 57.91% to 51.03% in i-FeV/Ti. As for p-FeV/Ti, it changes slightly from 63.57% to 60.26%. For K 2p XPS, the peaks located on the same place on both i-FeV/Ti and p-FeV/Ti (Fig. S9 in Supporting information).  $\text{K}_2\text{O}$  is a kind of electron donor that would damage the redox capacity of SCR catalysts, but the electron withdrawing function of Fe protect the true active sites of  $\text{VO}_x$  species from the attacking of  $\text{K}_2\text{O}$  [33]. As for the O 1s XPS spectra of all catalysts (Fig. S10 in Supporting information), the peaks located at  $\sim 531.46$  and  $\sim 530.16\text{ eV}$  are attributed to the surface absorbed oxygen (denoted as  $\text{O}_\alpha$ ) and the lattice oxygen (denoted as  $\text{O}_\beta$ ), respectively [45]. Generally,  $\text{O}_\alpha$  is easier to be



**Fig. 3.** (A) XPS spectra of V 2p and (B) Fe 2p of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti and K-p-FeV/Ti catalysts. (C, D)  $H_2$  temperature-programmed reduction ( $H_2$ -TPR) profiles and (E)  $NH_3$ -TPD-MS profiles of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti, and K-p-FeV/Ti catalysts. (F) *In situ* DRIFTS of  $NH_3$  desorption of i-FeV/Ti and p-FeV/Ti catalysts.

involved in the SCR reaction [46]. The  $O_{\alpha}/(O_{\alpha} + O_{\beta})$  ratios of i-FeV/Ti, K-i-FeV/Ti, p-FeV/Ti, K-p-FeV/Ti are 19.98%, 15.99%, 12.68%, and 7.61%, respectively. The higher amounts of  $O_{\alpha}$  on the surface of i-FeV/Ti can be resulted from the more O vacancies on the surface of i-FeV/Ti, which may be resulted from the strong interaction between Fe and V species. The introduction of  $K_2O$  would certainly cause the loss of O vacancies, leading to the decline of  $O_{\alpha}$  ratios [22]. While, the loss of O vacancies among K-i-FeV/Ti is less than that of K-p-FeV/Ti, further confirming the strong interaction between Fe and V species as well as the electron withdrawing function of Fe protect the active  $VO_x$  sites from  $K_2O$  poisoning.

In  $NH_3$ -SCR reaction, the acidity and redox properties of the catalyst are critical to its performance. The redox ability of catalysts was tested by  $H_2$  temperature-programmed reduction ( $H_2$ -TPR). As illustrated in Fig. 3C, the reduction peaks at 346, 372, and 387 °C for the i-FeV/Ti catalyst should be assigned to the reduction of  $Fe^{3+}$  to  $Fe^{(3-\delta)+}$ ,  $Fe^{(3-\delta)+}$  to  $Fe^{2+}$ , and  $V^{5+}$  to  $V^{4+}$ , respectively [22]. After  $K_2O$  poisoning, these peaks migrate to higher temperature and the reduction peak of  $Fe^{3+}$  to  $Fe^{(3-\delta)+}$ ,  $Fe^{(3-\delta)+}$  to  $Fe^{2+}$  overlap, indicating that the redox capacity of catalysts are affected. Besides, the peak attributed to  $V^{5+}$  to  $V^{4+}$  shifts from 387 °C to 414 °C after  $K_2O$  poisoning. Compared with p-FeV/Ti, it can be found that the redox capacity of V species on i-FeV/Ti is less affected. For p-FeV/Ti shown in Fig. 3D, the peaks located at 543 °C should be ascribed to  $V^{4+}$  to  $V^{3+}$ , which is not apparent on i-FeV/Ti, further evidencing the presence of more  $V^{4+}$  on the surface of p-FeV/Ti [47]. Furthermore, this peak becomes more pronounced after  $K_2O$  poisoning, indicating that  $K_2O$  poisoning exerts a greater influence on the V species of p-FeV/Ti. Besides, the  $H_2$  consumption of i-FeV/Ti (1.252 mmol/g) and K-i-FeV/Ti (1.375 mmol/g) are higher than p-FeV/Ti (0.909 mmol/g) and K-p-FeV/Ti (0.986 mmol/g), further proving the superior redox ability of i-FeV/Ti.

In addition to redox capacity, the acidity is also an important factor affecting the performance of SCR catalysts. Therefore,

the  $NH_3$  temperature-programmed desorption mass spectrometry ( $NH_3$ -TPD-MS) was undertaken to examine the acidity of all catalysts. As shown in Fig. 3E, the peak at 191 °C assigned to weak acid, the peak at 238 °C attributed to medium strong acid, and the other one at 318 °C presenting strong acid are observed on i-FeV/Ti. The  $NH_3$  desorption amount of i-FeV/Ti (19.19  $\mu\text{mol/g}$ ) is much higher than that of p-FeV/Ti (12.19  $\mu\text{mol/g}$ ). Actually, the  $NH_3$  adsorption ability of Fe species is quite poor [48], but the strong electron-withdrawing property enhances the acidity of  $VO_x$  species, which endows i-FeV/Ti with better  $NH_3$  adsorption. After  $K_2O$  poisoning, the  $NH_3$  desorption amount of K-i-FeV/Ti (7.38  $\mu\text{mol/g}$ ) remains higher than that of K-p-FeV/Ti (5.21  $\mu\text{mol/g}$ ). *In situ* diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) of  $NH_3$  desorption was also used to analyze the adsorption and desorption behavior of  $NH_3$  as well as the specific desorbed  $NH_3$  species in Fig. 3F. The bands at 1665, 1437  $\text{cm}^{-1}$  are ascribed to  $NH_4^+$  adsorbed on Brønsted acid sites, and bands located at 1602, 1235, 1183  $\text{cm}^{-1}$  are attributed to  $NH_3$  adsorbed on Lewis acid sites [49–55]. It can be observed that the intensity of the desorption peaks of i-FeV/Ti are significantly higher than those of p-FeV/Ti. These results demonstrate that i-FeV/Ti possess much stronger surface acidity than that of p-FeV/Ti. When the  $K_2O$  was incorporated, the adsorbed species as well as the peak intensities on i-FeV/Ti barely change at low temperatures (<200 °C) (Fig. S11 in Supporting information). As for p-FeV/Ti, there are dramatic reductions in peak intensities (Fig. S12 in Supporting information). The effect of  $K_2O$  on the surface acidity can be divided into two aspects: (1) occupying the Brønsted acid sites and (2) reducing the elemental valence thus weakening the Lewis acid strength through electronic effects. Therefore, due to the offset effect of the electron interactions caused by strong interaction of Fe and V on i-FeV/Ti, the Lewis acidity of active  $VO_x$  species is maintained after  $K_2O$  poisoning.

Furthermore, *in situ* DRIFTS were also displayed to clarify the  $NO$  adsorption behaviors over fresh and  $K_2O$ -poisoned catalysts.

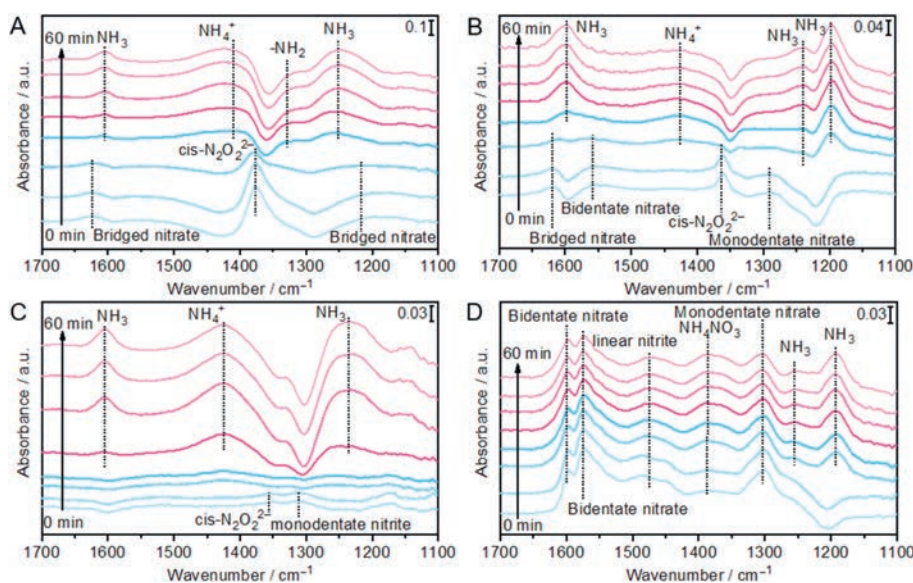


Fig. 4. *In situ* DRIFTS of the transient reactions at 240 °C between  $\text{NH}_3$  and preadsorbed  $\text{NO}+\text{O}_2$  of (A) i-FeV/Ti, (B) p-FeV/Ti, (C) K-i-FeV(p)/Ti, and (D) K-p-FeV/Ti catalysts.

For i-FeV/Ti, the bands at 1625, 1577, 1364 and 1294  $\text{cm}^{-1}$  are attributed to bridged nitrate, bidentate nitrate,  $\text{cis-N}_2\text{O}_2^{2-}$  and monodentate nitrate, respectively (Fig. S13 in Supporting information) [50,51,53,55]. As the temperature raised, the peaks at 1625 and 1577  $\text{cm}^{-1}$  gradually decline and those at 1364  $\text{cm}^{-1}$  increase, indicating that bridged nitrate and bidentate nitrate are progressively transformed to  $\text{cis-N}_2\text{O}_2^{2-}$ . It is reported that  $\text{cis-N}_2\text{O}_2^{2-}$  generated on the surface would greatly enhance the NO conversion [56]. That should be one reason why i-FeV/Ti exhibits superior SCR activity. As for p-FeV/Ti, the peaks at 1241 and 1623  $\text{cm}^{-1}$  ascribed to bridged nitrate are more stable (Fig. S14 in Supporting information) [51,52]. Compared with i-FeV/Ti, the peaks attributed to  $\text{cis-N}_2\text{O}_2^{2-}$  are weaker and appear at higher temperature, demonstrating its inferior redox ability. What is more, for p-FeV/Ti,  $\text{K}_2\text{O}$  addition promoted the formation of thermally stable absorbed  $\text{NO}_x$  species, which would cover the active sites, including bridged nitrate (1605  $\text{cm}^{-1}$ ), bidentate nitrate (1576  $\text{cm}^{-1}$ ), and monodentate nitrate (1300  $\text{cm}^{-1}$ ), leading to inferior  $\text{K}_2\text{O}$ -resistance (Fig. S15 in Supporting information) [57,58]. However, the intensities of absorbed nitrate on i-FeV/Ti are significantly reduced after  $\text{K}_2\text{O}$  poisoning (Fig. S16 in Supporting information). The negative effect of  $\text{K}_2\text{O}$  is greatly alleviated by the strong electronic interaction between Fe and V over i-FeV/Ti.

In order to clarify the whole reaction pathways over all catalysts, *in situ* DRIFTS of the transient reaction between  $\text{NO}+\text{O}_2$  and preadsorbed  $\text{NH}_3$  species were carried out at 240 °C, at which the activity difference between K-i-FeV/Ti and K-p-FeV/Ti is greatest. After  $\text{NH}_3$  is adsorbed on i-FeV/Ti for an hour, the surface absorbed ammonia species mainly include ionic  $\text{NH}_4^+$  (1421  $\text{cm}^{-1}$ ), coordinated  $\text{NH}_3$  (1601, 1255  $\text{cm}^{-1}$ ) and  $-\text{NH}_2$  amide species (1326  $\text{cm}^{-1}$ ) (Fig. S17 in Supporting information) [53,59–61]. They are consumed in 2 min with the addition of  $\text{NO}+\text{O}_2$ . Then the peaks ascribed to bridged nitrate (1619  $\text{cm}^{-1}$ ), bidentate nitrate (1256  $\text{cm}^{-1}$ ) and  $\text{cis-N}_2\text{O}_2^{2-}$  (1376  $\text{cm}^{-1}$ ) appear [57,62,63]. It is obvious that the absorbed ammonia species would easily participate in the SCR reaction. As for p-FeV/Ti, the peak intensities of ammonia species adsorbed on the surface, which should be attributed to  $\text{NH}_3$  (1602, 1243, and 1204  $\text{cm}^{-1}$ ) on the Lewis acid sites, are relatively lower, proving its inferior acidity (Fig. S18 in Supporting information) [51,52,55]. For the transient reaction at 240 °C between  $\text{NH}_3$  and preadsorbed  $\text{NO}+\text{O}_2$  in Fig. 4A,  $\text{cis-N}_2\text{O}_2^{2-}$  (1376  $\text{cm}^{-1}$ ) and bridged nitrate (1623, 1216  $\text{cm}^{-1}$ ) are

formed on i-FeV/Ti after adsorption of  $\text{NO}+\text{O}_2$  [51,57,63]. After the introduction of  $\text{NH}_3$ , these peaks vanish within two minutes and the bands belong to  $\text{NH}_4^+$  (1421  $\text{cm}^{-1}$ ),  $\text{NH}_3$  (1605, 1252  $\text{cm}^{-1}$ ) and  $-\text{NH}_2$  amide species (1329  $\text{cm}^{-1}$ ) appear, proving that these nitrates possess high reactivity with  $\text{NH}_3$  [53,55,60,61]. For p-FeV/Ti in Fig. 4B, after injecting  $\text{NO}+\text{O}_2$  for 1 h, the peaks attributed to bridged nitrate (1619  $\text{cm}^{-1}$ ), bidentate nitrate (1558  $\text{cm}^{-1}$ ),  $\text{cis-N}_2\text{O}_2^{2-}$  (1362  $\text{cm}^{-1}$ ), and monodentate nitrate (1291  $\text{cm}^{-1}$ ) form and these adsorbed nitrate species could contribute to the SCR reaction [50,53,61]. However, the intensity of these peaks is quite lower than those observed in i-FeV/Ti, indicating its inferior redox ability. According to the above phenomena, we could find that the preadsorbed  $\text{NH}_3$  species could react with  $\text{NO}+\text{O}_2$  and the preadsorbed  $\text{NO}_x$  species could also react with  $\text{NH}_3$  on these fresh catalysts, demonstrating that the reactions on the surface would both follow Langmuir-Hinshelwood mechanism and Eley-Rideal mechanism.

To elucidate the effect of  $\text{K}_2\text{O}$  addition on the reaction pathway, *in situ* DRIFTS of the transient reactions were also carried out on  $\text{K}_2\text{O}$  poisoned catalysts. Compared with fresh i-FeV/Ti, similar species such as  $\text{NH}_3$  (1605, 1240  $\text{cm}^{-1}$ ),  $\text{NH}_4^+$  (1425  $\text{cm}^{-1}$ ) and  $-\text{NH}_2$  (1332  $\text{cm}^{-1}$ ) form on K-i-FeV/Ti after the injection of  $\text{NH}_3$  for 1 h (Fig. S19 in Supporting information), but the intensities of these bands are reduced, which indicate that the addition of  $\text{K}_2\text{O}$  suppress the acidity on the surface [49,53,55,61]. A similar phenomenon also occurs over K-p-FeV/Ti, but the differences are that the peak intensities are quite lower on K-p-FeV/Ti. After the introduction of  $\text{NO}+\text{O}_2$ , the adsorbed  $\text{NH}_3$  species are quickly consumed and the subsequent peaks attributed to absorbed  $\text{NO}_x$  species are quite weak on i-FeV/Ti, while a large amount of inactive nitrate species such as bidentate nitrate (1597, 1574  $\text{cm}^{-1}$ ), linear nitrite (1484  $\text{cm}^{-1}$ ) and monodentate nitrate (1300  $\text{cm}^{-1}$ ) accumulate on K-p-FeV/Ti (Fig. S20 in Supporting information) [52,55,60,61]. Similar tendency also occurs on the transient reaction at 240 °C between  $\text{NH}_3$  and preadsorbed  $\text{NO}+\text{O}_2$ . After preadsorption of  $\text{NO}+\text{O}_2$  for 1 h, only quite few amounts of  $\text{NO}_x$  species such as  $\text{cis-N}_2\text{O}_2^{2-}$  (1356  $\text{cm}^{-1}$ ), monodentate nitrite (1311  $\text{cm}^{-1}$ ) and chelating nitrite (1174  $\text{cm}^{-1}$ ) are preserved on i-FeV/Ti in Fig. 4C. With the addition of  $\text{NH}_3$ , they are consumed in 5 min and corresponding  $\text{NH}_3/\text{NH}_4^+$  species form on the surface [50,55,58]. In contrast, after the flowing of  $\text{NO}+\text{O}_2$  for 1 h, a large number of stable nitrate species cover the surface of p-FeV/Ti in Fig. 4D. It was reported that K

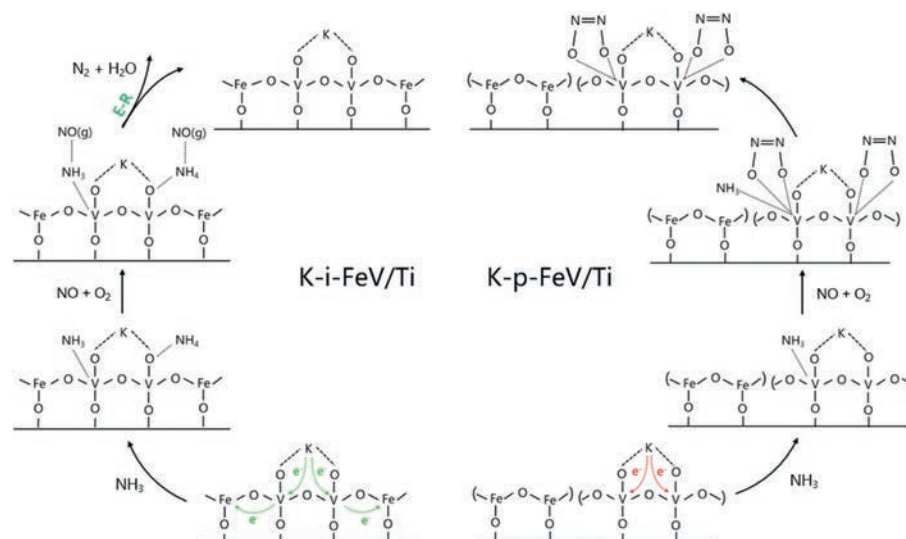


Fig. 5. Possible reaction pathway of K poisoned i-FeV/Ti and p-FeV/Ti catalysts.

would not affect the formation of ammonium nitrate species, but it would restrain the decomposition of ammonium nitrate species to  $N_2$  and  $H_2O$  [28]. With the introduction of  $NH_3$ , these peaks do not change at all while some bands ( $1254\text{ cm}^{-1}$ ,  $1193\text{ cm}^{-1}$ ) attributed to the adsorbed  $NH_3$  appear. It indicates that these stable nitrates cover the active V sites and hinder the process of SCR reaction on K-p-FeV/Ti, while on K-i-FeV/Ti, the strong electron interaction between Fe and V protect the active V species and promote SCR reaction [55,60]. Based on these results, after K poisoning, the reactions occur on i-FeV/Ti mainly follow the Eley-Rideal mechanism while the Langmuir-Hinshelwood mechanism is weakened, which is similar with other K poisoned V-based catalysts [11].

In summary, the effects of electron interactions on alkali resistance over  $FeVO_4/TiO_2$  for  $NH_3$ -SCR are revealed. From Fig. 5, it can be found that due to the strong electron-withdrawing effect of Fe, the valence of V element on the surface is increased, which facilitates the adsorption and activation of  $NH_3$ . After alkali poisoning, the electron-absorbing effect of Fe offsets the electron-giving effect of  $K_2O$  on V element, thus protecting the real active sites of  $VO_x$  species. It allows the catalysts to participate in the SCR reaction through E-R mechanism even in the presence of  $K_2O$ . This work provides a profound comprehension of the electron interactions effects on the alkali resistance of catalysts and paves a promising way for the design of highly alkali-resistant SCR catalysts in the future.

### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

### Acknowledgments

We acknowledge the National Natural Science Foundation of China (No. 22125604), Shanghai Rising-Star Program (No. 22QA1403700) and Chenguang Program supported by Shanghai Education Development Foundation and Shanghai Municipal Education Commission (No. 22Z00354).

### Supplementary materials

Supplementary material associated with this article can be found, in the online version, at doi:10.1016/j.ccl.2023.108635.

### References

- [1] S. Beirle, K.F. Boersma, U. Platt, et al., *Science* 333 (2011) 1737–1739.
- [2] L. Han, S. Cai, M. Gao, et al., *Chem. Rev.* 119 (2019) 10916–10976.
- [3] H. Wang, T. Zhu, Y. Qiao, et al., *Chin. Chem. Lett.* 33 (2022) 5223–5227.
- [4] A. Marberger, D. Ferri, M. Elsener, et al., *Angew. Chem. Int. Ed.* 55 (2016) 11989–11994.
- [5] W. Song, J. Ji, K. Guo, et al., *Chin. Chem. Lett.* 33 (2022) 935–938.
- [6] G. He, Z. Lian, Y. Yu, et al., *Sci. Adv.* 4 (2018) eaau4637.
- [7] N.R. Jaegers, J.K. Lai, Y. He, et al., *Angew. Chem. Int. Ed.* 58 (2019) 12609–12616.
- [8] J. Ji, L. Han, W. Song, et al., *Chin. Chem. Lett.* 23 (2022) 107769.
- [9] X. Liu, P. Wang, Y. Shen, et al., *Environ. Sci. Technol.* 56 (2022) 11646–11656.
- [10] Y. Pu, X. Xie, W. Jiang, et al., *Chin. Chem. Lett.* 31 (2020) 2549–2555.
- [11] G. Zhou, P. Maitarad, P. Wang, et al., *Environ. Sci. Technol.* 54 (2020) 13314–13321.
- [12] F. Tang, B. Xu, H. Shi, et al., *Appl. Catal. B* 94 (2010) 71–76.
- [13] Z. Shen, X. Liu, S. Impeng, et al., *Environ. Sci. Technol.* 56 (2022) 5141–5149.
- [14] Z. Jia, Y. Shen, T. Yan, et al., *Environ. Sci. Technol.* 56 (2022) 2647–2655.
- [15] Y. Zhao, L. Shi, Y. Shen, et al., *Environ. Sci. Technol.* 56 (2022) 4386–4395.
- [16] Z. Liu, S. Zhang, J. Li, et al., *Appl. Catal. B* 158–159 (2014) 11–19.
- [17] N. Zhu, W. Shan, Z. Lian, et al., *J. Hazard. Mater.* 382 (2020) 120970.
- [18] G. Wu, J. Li, Z. Fang, et al., *Chem. Eng. J.* 271 (2015) 1–13.
- [19] S. Li, W. Hu, Z. Xu, et al., *ChemCatChem* 14 (2022) e202200476.
- [20] F. Liu, H. He, Z. Lian, et al., *J. Catal.* 307 (2013) 340–351.
- [21] J. Mu, X. Li, W. Sun, et al., *ACS Catal.* 8 (2018) 6760–6774.
- [22] Z. Si, Y. Shen, J. He, et al., *Environ. Sci. Technol.* 56 (2022) 605–613.
- [23] Y. Li, S. Cai, P. Wang, et al., *Environ. Sci. Technol.* 55 (2021) 9276–9284.
- [24] P. Zhang, P. Wang, A. Chen, et al., *Environ. Sci. Technol.* 55 (2021) 11970–11978.
- [25] C. Feng, P. Wang, X. Liu, et al., *Environ. Sci. Technol.* 55 (2021) 11255–11264.
- [26] K. Zha, C. Feng, L. Han, et al., *Chem. Eng. J.* 381 (2020) 122764.
- [27] Z. Zhou, J. Lan, L. Liu, et al., *Catal. Commun.* 149 (2021) 106230.
- [28] S. Xiong, J. Chen, H. Liu, et al., *Environ. Sci. Technol.* 56 (2022) 3739–3747.
- [29] Y. Zheng, Y. Guo, W. Shan, et al., *Appl. Catal. A* 637 (2022) 118606.
- [30] H. Li, J. Miao, Q. Su, et al., *J. Mater. Sci.* 54 (2019) 14707–14719.
- [31] M. Casanova, L. Nodari, A. Sagar, et al., *Appl. Catal. B* 176–177 (2015) 699–708.
- [32] S. Li, W. Huang, H. Xu, et al., *Appl. Catal. B* 270 (2020) 118872.
- [33] S. Zhang, S. Liu, W. Hu, et al., *Appl. Surf. Sci.* 466 (2019) 99–109.
- [34] M. Zhu, J.K. Lai, U. Tumuluri, et al., *ACS Catal.* 7 (2017) 8358–8361.
- [35] H. Liu, J. Wang, T. Yu, et al., *Catal. Sci. Technol.* 4 (2014) 1350–1356.
- [36] F. Gao, Y. Zheng, R.K. Kukkadapu, et al., *ACS Catal.* 6 (2016) 2939–2954.
- [37] N. Zhu, Z. Lian, Y. Zhang, et al., *Chin. Chem. Lett.* 30 (2019) 867–870.
- [38] L.J. Burcham, G. Deo, X.T. Gao, et al., *Top. Catal.* 11 (2000) 85–100.
- [39] A. Marberger, M. Elsener, D. Ferri, et al., *ACS Catal.* 5 (2015) 4180–4188.
- [40] Z. Lian, J. Wei, W. Shan, et al., *J. Am. Chem. Soc.* 143 (2021) 10454–10461.
- [41] K. Routray, W. Zhou, C.J. Kiely, et al., *ACS Catal.* 1 (2010) 54–66.
- [42] H. Tian, I.E. Wachs, L.E. Briand, *J. Phys. Chem. B* 109 (2005) 23491–23499.
- [43] B. Liu, X. Li, Q. Zhao, et al., *J. Mater. Chem. A* 3 (2015) 15163–15170.
- [44] T.H. Vuong, J. Radnik, J. Rabeah, et al., *ACS Catal.* 7 (2017) 1693–1705.
- [45] L. Kang, L. Han, P. Wang, et al., *Environ. Sci. Technol.* 54 (2020) 14066–14075.
- [46] L. Yan, F. Wang, P. Wang, et al., *Environ. Sci. Technol.* 54 (2020) 7697–7705.
- [47] G. Wu, J. Li, Z. Fang, et al., *Catal. Commun.* 64 (2015) 75–79.
- [48] H. Wang, Z. Qu, S. Dong, et al., *Environ. Sci. Technol.* 50 (2016) 13511–13519.
- [49] W. Chen, Z. Li, F. Hu, et al., *Appl. Surf. Sci.* 439 (2018) 75–81.
- [50] Y. Zhang, X. Yue, T. Huang, et al., *Materials* 11 (2018) 1307.
- [51] Q. Zhang, J. Fan, P. Ning, et al., *Appl. Surf. Sci.* 435 (2018) 1037–1045.
- [52] L. Huang, K. Zha, S. Namuangruk, et al., *Catal. Sci. Technol.* 6 (2016) 8516–8524.

- [53] Z. Liu, S. Zhang, J. Li, et al., *Appl. Catal. B* 144 (2014) 90–95.
- [54] L. Wei, S. Cui, H. Guo, et al., *J. Mol. Catal. A: Chem.* 421 (2016) 102–108.
- [55] H. Hu, S. Cai, H. Li, et al., *ACS Catal.* 5 (2015) 6069–6077.
- [56] K. Cheng, B. Liu, W. Song, et al., *Ind. Eng. Chem. Res.* 57 (2018) 7802–7810.
- [57] H. Liang, K. Gui, X. Zha, *Can. J. Chem. Eng.* 94 (2016) 1668–1675.
- [58] W. Mu, J. Zhu, S. Zhang, et al., *Catal. Sci. Technol.* 6 (2016) 7532–7548.
- [59] J. Liu, J. Meeprasert, S. Namuangruk, et al., *J. Phys. Chem. C* 121 (2017) 4970–4979.
- [60] Y. Yu, J. Wang, J. Chen, et al., *Ind. Eng. Chem. Res.* 53 (2014) 16229–16234.
- [61] Z. Ma, X. Wu, H. Härelind, et al., *J. Mol. Catal. A: Chem.* 423 (2016) 172–180.
- [62] M.M. Azis, H. Härelind, D. Creaser, *Catal. Sci. Technol.* 5 (2015) 296–309.
- [63] Z. Zhang, L. Chen, Z. Li, et al., *Catal. Sci. Technol.* 6 (2016) 7151–7162.